

1 Title

2 How the relationship between vegetation cover and land-cover variance constrains
3 biodiversity in a human dominated world

4 Authors

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16 Abstract

17 Context: Alteration of natural vegetation cover across the landscape drives biodiversity
18 changes. Although several studies have explored the relationships between vegetation

19 cover and species richness, as well as between land-cover variance and species richness,
20 few have considered the non-independence of these two biodiversity drivers.

21 Objectives: The goal of this perspective paper is to present theoretical and empirical
22 relationships linking vegetation cover to land-cover variance at the landscape scale, and
23 the implication of these relationships for species richness change along a gradient of
24 increasing anthropization.

25 Methods and results: We used simulated and empirical Normalized Difference
26 Vegetation Index data to examine the generality of the relationship between vegetation
27 cover and land-cover variance. Using the province of Québec (Canada) as a case study,
28 our results show that decreasing vegetation cover captures the transition from
29 landscapes with low land-cover variance (non-anthropized landscapes), to intermediate
30 variance (agricultural landscapes), to high variance (urban landscapes).

31 Conclusion: Based on this relationship between vegetation cover and land-cover
32 variance, and assuming independent positive monotonic relationships between
33 biodiversity and both of these drivers, we predict a unimodal relationship between
34 species richness and anthropization. This suggests a threshold of anthropization beyond
35 which the positive effects of land-cover variance no longer compensate for the negative
36 effects of vegetation cover loss. Identifying these thresholds could be key to setting
37 conservation targets at a landscape scale.

38 **Keywords**

39 biodiversity, conservation, land-cover variance, species richness, vegetation cover,
40 environmental heterogeneity, landscape composition, landscape structure

41

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48 Introduction

49 A central goal of conservation ecology is to understand how organisms respond to our
50 increasing human footprint. Organisms' responses to habitat alteration are highly
51 complex, multifaceted, and variable among species and across contexts (Bender et al.
52 1998; Debinski and Holt 2000; Connor et al. 2000; Prugh et al. 2008). Nevertheless, at
53 a landscape scale, the net effect of anthropization is driven in large part by the
54 response of species to changes in vegetation cover and/or changes in spatial land-cover
55 variance. Early conceptual models in landscape ecology expressed the anthropization
56 gradient as follows (Forman 1995; McIntyre and Hobbs 1999): 1) start with a uniform
57 natural landscape with relatively high vegetation cover and low land-cover variance;
58 2) progress to a heterogeneous semi-natural landscape with an intermediate vegetation
59 cover and increased land-cover variance; 3) end with an impacted landscape with
60 relatively low vegetation cover and low land-cover variance. This trajectory suggests a
61 decreasing relationship between the anthropization gradient (i.e., transitioning from
62 stages 1 to 3 above) and vegetation cover, and a hump-shaped relationship between the
63 anthropization gradient and land-cover variance.

64

65 We define *vegetation cover* as the average leaf area per unit of land area, typically
66 known as the Leaf Area Index. Vegetation cover changes over space in relation to the
67 proportion of land occupied by plants, but also how tall or densely packed plants are.
68 In practice, vegetation cover can be quantified using surrogate measures like the
69 normalized difference vegetation index (NDVI) or the enhanced vegetation index
70 (EVI). Mean vegetation cover decreases along an anthropization gradient if newly
71 created land-cover patches, such as roads, abandoned fields, or crops, do not contain

72 as much leaf area as the original vegetation. Likewise, we define land-cover variance
73 as the magnitude of contrast (in terms of vegetation cover) among different areas
74 within the landscape. Land-cover variance thus measures only compositional
75 heterogeneity, and ignores its configurational component (*sensu* Fahrig et al. 2011).
76 For present purposes, we quantified vegetation cover as the mean of NDVI values
77 across a landscape and land-cover variance as the mathematical variance of these
78 NDVI values.

79

80 Although vegetation cover and land-cover variance are well-studied variables, few
81 authors have investigated how the two are interrelated. We expect they are related, as
82 this is often the case for the mean and variance of measured variables (here, NDVI)
83 (e.g., Tokeshi, 1995). Possible relations for vegetation cover and land-cover variance
84 have been postulated in conceptual models (Tilman and Pacala 1993; Abrams 1995),
85 but the relationship between them remains unexplored. In this perspective paper, we
86 present both theoretical and empirical arguments for the presence of a non-linear
87 hump-shaped relationship between vegetation cover and land-cover variance at
88 landscape scales. We discuss the ecological and conservation implications of this
89 relationship, providing a new perspective on productivity-biodiversity and
90 heterogeneity-biodiversity relationships observed in nature.

91 How are vegetation cover and land-cover variance related?

92 A mathematical example

93 Let us imagine a landscape in which vegetation cover is measured on a continuous
94 scale, for example, as NDVI on the bounded interval $[-1,1]$. Each portion (pixel) of the
95 landscape has a value of NDVI. Let us now assume that the distribution of vegetation
96 cover values over the landscape is described statistically by the Beta distribution:

97
$$P(x) = \frac{\Gamma(\alpha+\beta)}{\Gamma(\alpha)\Gamma(\beta)} x^{\alpha-1}(1-x)^{\beta-1},$$

98 which is often the case for ecological variables with bounded distributions (Stoy et al.
99 2009). Following our definitions of vegetation cover and land-cover variance, the
100 mean is directly related to the α and β parameters of the Beta distribution as follows:

101
$$E[X] = \frac{\alpha}{\alpha + \beta},$$

102 and its variance depends on the same two parameters:

103
$$var[X] = \frac{\alpha \beta}{(\alpha + \beta)^2 (\alpha + \beta + 1)}$$

104 Hence, the mean vegetation cover of a landscape will directly affect the land-cover
105 variance that can be observed (Remmel 2009). The intuition behind such mathematical
106 statements is that, in the context of a bounded distribution, large deviations around the
107 mean can only exist when the mean is towards the middle of the range of possible
108 values. Any distribution with a mean near the upper or lower bound of the range
109 implies small deviations around that mean. This relationship can be visualized by
110 sampling evenly spaced intervals of α and β parameters and calculating the mean and
111 variance of each combination. Technically, the procedure is equivalent to simulating
112 landscapes over a large range of NDVI spatial patterns and calculating the mean and

113 variance of these values. Plotting the mean vs. variance of NDVI values in these
114 landscapes reveals a clear hump-shaped envelope between them (Fig. 1A). Each
115 combination of mean and variance in Fig. 1A translates into a different type of
116 landscape pattern.

117

118 Empirical illustration

119 To characterize the predicted mean-variance relationship in real landscapes, we
120 calculated vegetation cover and land-cover variance using Sentinel-2 radiometric data
121 (Drusch et al. 2012). We selected 10 000 random longitude and latitude coordinates
122 from a gaussian distribution centered around Montréal, QC (Latitude 45.50, Longitude
123 -73.56, standard deviation 2.5 degrees) covering the whole gradient of ecosystems
124 found throughout the province of Québec (Canada), including tundra, boreal, mixed,
125 and deciduous forests, as well as urban and agricultural areas. A land-cover
126 classification conducted using 2015 data from the Operational Land Imager (OLI)
127 Landsat sensor (Government of Canada et al. 2019) was associated with each pair of
128 coordinates. We removed coordinates falling outside the boundaries of Québec or
129 directly on water, ice, snow or wetland land-cover categories, leaving 8 970
130 landscapes.

131

132 Gradients in vegetation cover and land-cover variance in this dataset stem from a
133 combination of natural (e.g., across biomes) and anthropized areas. For each pair of
134 coordinates, we used Google Earth Engine (Gorelick et al. 2017) to create circular
135 100 m radius and 564 m radius (i.e., 1 km²) landscapes centred on the coordinates on
136 Sentinel-2 images (10 m ground resolution). These sizes correspond respectively to the

137 radius around which bird songs can typically be heard in point count surveys and the
138 home range size of most passerine birds (Brown and Sullivan 2005). We refer to these
139 arbitrarily-selected scales as small and large landscapes hereafter. We selected all
140 Sentinel-2 images taken between 1 Jun and 1 Sep 2018, to ensure that they are
141 representative of summertime vegetation cover. For each band and each pixel forming
142 the image, we calculated the median value of the time series to remove the influence
143 of cloud cover and cloud shadows (Namikawa 2017). We then calculated the
144 Normalized Difference Vegetation Index (NDVI; Rouse Jr et al. 1974) for each pixel,
145 using bands B8 (near infrared) and B4 (red). In a subsequent step, we extracted NDVI
146 mean and variance across pixels, which we interpreted as vegetation cover and land-
147 cover variance values for each landscape. Graphically representing these variables
148 together revealed a hump shape, which we illustrated using a thin-plate spline with
149 penalized regression coefficients (Wood 2003). Spline functions fitted to small (100 m
150 radius) and large landscapes (564 m radius) explained respectively 28 and 41% of
151 land-cover variance using only vegetation cover as the explanatory variable (Fig 1B
152 and 1C).

153

154 The vegetation cover gradient captures the transition from natural landscapes with a
155 high vegetation cover (low land-cover variance), to croplands (moderate variance), to
156 urban lands (high variance), and finally to landscapes with a naturally low vegetation
157 cover (low variance) (Fig. 1). In particular, several urban landscapes in southern
158 Québec show very high levels of land-cover variance; i.e., black points in the upper
159 left of the envelope (Fig. 1B and 1C). Considering the proximity of these landscapes
160 to the upper boundary of the relationship, further losses of vegetation cover are likely
161 to translate into decreasing land-cover variance. In other words, the land-cover

162 variance of urban landscapes at the boundary will start decreasing if they are pushed
163 farther to the left on the vegetation cover axis (Fig. 1A).

164

165 The extent of the landscape influences some aspects of the cover-variance relationship.
166 Although the overall shape of the data envelope did not change, the strength of the
167 relationship increased with increasing spatial extent. Indeed, at a large spatial extent it
168 is hard to find landscapes with either very high or very low land-cover variance, thus
169 decreasing the variation of observed values on that axis. Spatial extent also influences
170 the relative position of landscapes along the vegetation cover axis. For instance, the
171 same urban area could reveal a low vegetation cover at a small spatial extent (e.g., a
172 parking lot), but a much higher vegetation cover if observed at a larger extent that
173 incorporates green spaces such as residential yards or municipal parks. The effect of
174 extent will be especially marked if the landscape overlays different land-cover
175 classifications (e.g., urban and cropland areas).

176

177 Implications for biodiversity conservation

178 The relationships between vegetation cover and land-cover variance described above
179 have implications for our understanding of how species richness changes along an
180 anthropization gradient. To highlight these implications, let us consider circumstances
181 where species richness increases monotonically with both vegetation cover and land-
182 cover variance. Such a scenario is supported by an extensive literature on the
183 productivity-species diversity hypothesis (Cusens et al. 2012; Gillman et al. 2015) and
184 the heterogeneity-species diversity hypothesis (Tews et al. 2004; Stein et al. 2014).

185 The above hypotheses are represented by positive monotonic relationships between
186 vegetation cover and species richness, as well as between land-cover variance (i.e.,
187 spatial heterogeneity) and species richness (Fig. 2).

188

189 Conceptually, the anthropization gradient captures how landscapes change as they
190 move along the non-linear relationship between vegetation cover and land-cover
191 variance, and combines their joint effects. These gradients are represented by a
192 decreasing relationship between anthropization and vegetation cover, and a hump-
193 shaped relationship between anthropization and land-cover variance (Fig 2). Assuming
194 for our purposes that vegetation cover and land-cover variance have independent and
195 additive effects on the species richness of a landscape, a concave relationship is then
196 obtained between species richness and the level of anthropization (Fig 2).

197

198 From our investigation, we speculate that confusion may arise when studying the
199 bivariate empirical relationship between biodiversity variables and vegetation cover at
200 the landscape scale. Although many species might respond monotonically to
201 vegetation cover (Gilroy et al. 2014), our results above reveal that vegetation cover
202 and land-cover variance are in fact interrelated in a non-linear fashion (Fig. 1). Thus,
203 the observed bivariate relationship between vegetation cover and species richness
204 could appear hump-shaped; although this would be the net result of the two factors
205 acting at one on species richness (see Evans et al. 2005).

206

207 Our conceptual model predicts a concave response of species richness to
208 anthropization whenever vegetation cover and land-cover variance are respectively a
209 monotonic decreasing and a unimodal function of landscape anthropization. Other

210 studies support the idea that species richness varies unimodally with anthropization.
211 The PREDICTS project, for Projecting Responses of Ecological Diversity In
212 Changing Terrestrial Systems, is a large concerted effort to better understand the
213 influence of land-use change on species richness and composition. The database
214 consists of 320,924 records at 11,525 sites, from 284 publications, including 26,953
215 species from 13 terrestrial biomes (Newbold et al. 2015). The main results of
216 PREDICTS show that intensive land-use of plantation, cropland, or urbanization is
217 associated with a steep decline of species richness based on rarefied estimates (ca. -
218 40%). In contrast, areas of lower human population density, as well as extensive land
219 cover of primary and secondary natural vegetation, are associated with a slight
220 increase in species richness (ca. +5%). While there exists considerable spatial
221 variation in the data, the overall inference would be that species richness should
222 increase, on average, with increasing human population from low to moderate density
223 (up to 20-40 persons/km²), then decrease steadily at higher human density (Newbold
224 et al. 2015). In particular, the relationship between species richness and human
225 population density (Extended data Figure 2 in Newbold et al. 2015) bears striking
226 similarities with the conceptual ones presented herein (Fig. 3).

227

228 Another global analysis of 375 studies distributed worldwide evaluated the effect of
229 land-use intensity on plant diversity (Gerstner et al. 2014). The study found negative
230 effects of intensive land-use practices, like nutrient-input farming and tree plantations,
231 on plant diversity (correlation of ca. -0.46). Yet, the study also reports positive,
232 although weak, effects on plant diversity of extensive management practices, such as
233 mowing/grazing of grasslands and logging/thinning of forests (correlation of ca.
234 +0.14). Remarkably, abandonment of extensively managed landscapes (e.g., meadow

235 or pasture) decreased plant diversity, while abandonment of intensively managed
236 landscapes (e.g., conventional farming or tree plantation) increased diversity (Gerstner
237 et al. 2014).

238

239 We emphasize that the relationship we describe here is not equivalent to other
240 conceptual models that have predicted a hump-shaped response of species richness to
241 increasing temporal disturbance (the intermediate disturbance hypothesis; Connell,
242 1978) or more recently to spatial heterogeneity; i.e. the intermediate heterogeneity
243 hypothesis (Fahrig et al. 2011) also called the area-heterogeneity trade-off (Allouche
244 et al. 2012). While the intermediate disturbance hypothesis was criticized on both
245 empirical and theoretical grounds (see Fox 2013), the area-heterogeneity trade-off has
246 received some support (Yang et al. 2015, Shuler et al. 2017, but see Ben-Hur and
247 Kadmon 2020). These two hypotheses suggest that “environmental variation”, either
248 in time or space, drives species coexistence. In contrast, we herein propose that this
249 environmental variation is a (non-linear) function of vegetation cover change, and that
250 both factors affect species richness at the landscape scale. We also note that our
251 framework has the advantage of clearly defining landscape properties in terms of mean
252 and variance of the vegetation cover. This last point is particularly important as
253 concepts like “disturbance” or “heterogeneity” have received multiple interpretation in
254 the literature (e.g., Stein and Kreft 2015). Even the term “area” in the area-
255 heterogeneity trade-off can be challenging in practice because it implies that species
256 habitats are discrete, measurable quantities. However, the generality of the framework
257 we propose also has its downsides, as it makes no prediction about which species
258 should be favoured or disfavoured with increasing anthropization. For example,
259 species richness may increase mostly through the addition of cosmopolitan exotic

260 species as well as a few disturbance-tolerant native species (McKinney and Lockwood
261 1999, McCune and Vellend 2013).

262 Future directions

263 In a metapopulation-based model describing species coexistence in patches of
264 fragmented habitats, Tilman et al. (1994) showed that the relationship between the
265 proportion of habitat destroyed and the proportion of species driven extinct should be
266 relatively flat in the early stages of anthropization and rapidly accelerate with further
267 habitat loss. Syntheses of biodiversity changes at landscape scales in the Anthropocene
268 point to a flat response, or even a slight increase with anthropization of species
269 richness in plants (Vellend et al. 2017) and vertebrates (Pautasso 2007). Such
270 compensatory effects could precede the steeper and more consequential part of Tilman
271 et al.'s (1994) prediction, where species extirpation rapidly increases following further
272 habitat losses. However, Tilman's model assumes that no species can survive in
273 habitats with a vegetation cover below some level. The conceptual model we propose
274 here offers a simple, yet realistic, explanation for the observed increase in biodiversity
275 with landscape anthropization. Increasing land-cover variance in the early stages of
276 anthropization might compensate for natural habitat loss (Fig. 1). It suggests that each
277 landscape has a threshold beyond which the loss of vegetation cover will switch from
278 increasing to decreasing variance, with synergistic negative effects on the persistence
279 of species. Identifying where this shift occurs along various anthropization gradients
280 and environmental contexts could be key in setting conservation targets at a landscape
281 scale. Future work on this topic should explore the influence of spatial scale (Fig 1), as

282 well as assumptions of a linear mapping of species richness on vegetation cover and
283 land-cover variance (Fig 2).

284

285 In summary, our framework implies that species richness responses to land-use
286 intensification result from the opposing effects of decreasing vegetation cover and
287 increasing land-cover variance at low to moderate anthropization levels, and from
288 synergistic negative effects of these two factors at high anthropization levels. Results
289 from two global syntheses so far corroborate these general predictions across many
290 species and environmental contexts (Gerstner et al. 2014; Newbold et al. 2015).

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299 **Consent to participate:** Not applicable

300 **Consent for publication:** All authors consent to publication.

301

302 **Availability of data and material:** The data are available on Figshare:

303 [10.6084/m9.figshare.12937241](https://doi.org/10.6084/m9.figshare.12937241)

304 **Code availability:** The code is available on Figshare: [10.6084/m9.figshare.12937241](https://doi.org/10.6084/m9.figshare.12937241)

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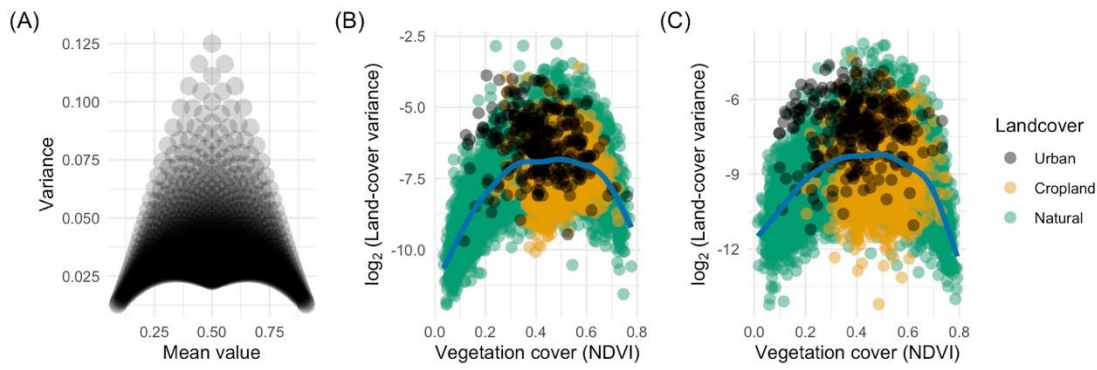
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416 **Figures**



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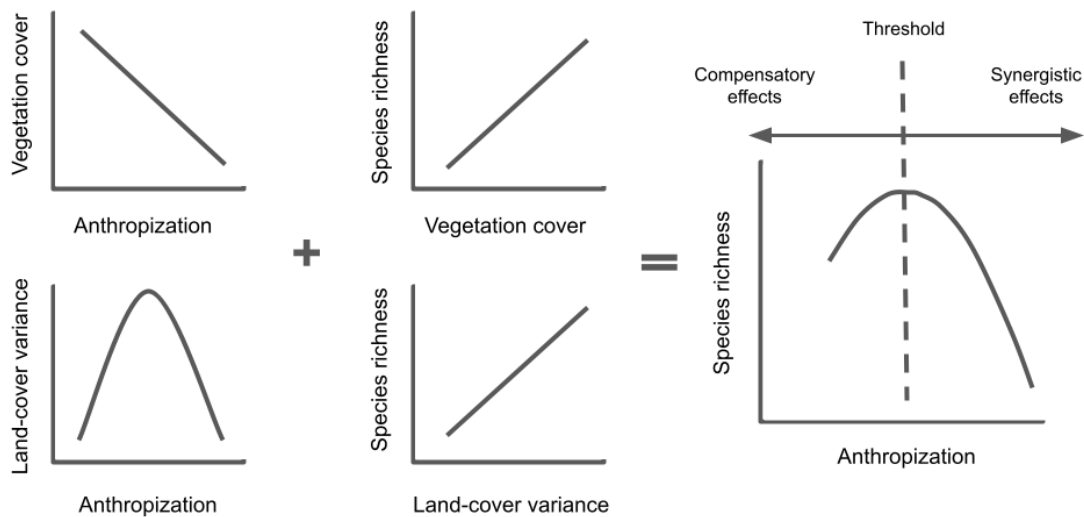
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Fig. 1 Mathematical and empirical relationship between vegetation cover and land-cover variance. Each point on a plot represents a prediction (A) or observation (B and C) for an individual landscape. (A) Predicted relationship between vegetation cover and land-cover variance based on the mathematical parameterization of a Beta distribution. Each pair of mean and variance values notionally represents the mean and variance of vegetation cover for a hypothetical landscape. They are sampled from the Beta distribution, using an evenly spaced grid of α and β parameters, both ranging from 0.5 to 5.0. (B and C) Measured relationship between vegetation cover and land-cover variance measured in 8970 landscapes across the province of Quebec, Canada, at two different scales (B: 564-m-radius landscapes, C: 100-m-radius landscapes) with corresponding land-use category. Vegetation cover and land use variance were calculated for NDVI values across each landscape. Land-cover variance was log-2 transformed to better visualize the hump-shaped pattern.



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433 Fig 2. Conceptual representation of the net effect of vegetation cover and land-cover
 434 variance on species richness along the anthropization gradient. The framework
 435 assumes positive monotonous relationships between vegetation cover and species
 436 richness, as well as between land-cover variance (i.e., spatial heterogeneity) and
 437 species richness. Conceptually, the anthropization gradient captures how landscapes
 438 change as they move along the non-linear relationship between vegetation cover and
 439 land-cover variance, and combines their joint effects. These gradients are represented
 440 by a decreasing relationship between anthropization and vegetation cover, and a
 441 hump-shaped relationship between anthropization and land-cover variance.
 442 Assuming that vegetation cover and land-cover variance have independent and
 443 additive effects on the species richness of a landscape, a concave relationship is
 444 obtained between species richness and the level of anthropization.