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ÉVALUATION ISOTOPIQUE DE L'IMPORTANCE RELATIVE DES  
PRODUCTIONS PHYTOPLANCTONIQUE VS PÉRIPHYTONIQUE DES ZONES  
DE FAIBLES VÉLOCITÉS POUR LES CONSOMMATEURS PRIMAIRES D'UN  
ÉCOSYSTÈME FLUVIAL

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## AVANT-PROPOS

En accord avec les articles 136 et 138 du *Règlement des études de cycles supérieurs* de l'UQTR, le présent document est présenté sous la forme d'un article scientifique. Il est divisé en deux chapitres, un premier présentant un résumé substantiel (en français) du mémoire et un second présentant l'article (en anglais) qui sera soumis à la revue *Limnology and Oceanography* et intitulé *Isotopic evaluation of the relative importance of planktonic vs periphytic production in a large river's slackwater for primary consumers*.

## RÉSUMÉ

Les isotopes stables du carbone ( $\delta^{13}\text{C}$ ) sont des traceurs naturels de plus en plus utilisés pour estimer la contribution relative des algues périphytiques et planctoniques à la production secondaire des rivières (invertébrés, poissons). Cette approche, essentiellement développée dans des études de lacs profonds et stratifiés, reste à être vérifiée dans le contexte des zones de faibles vitesses des écosystèmes fluviaux, où le temps de résidence est court, la stratification est faible, la turbidité variable et les patrons de mélange complexes, dû à l'apport des tributaires. Nous avons analysé des filtreurs et des brouteurs, à 12 stations dans un lac fluvial du fleuve Saint-Laurent, Canada. Nous avons observé une forte variation du  $\Delta\delta^{13}\text{C}$  ( $\delta^{13}\text{C}$  des brouteurs –  $\delta^{13}\text{C}$  des filtreurs), allant de 1 à 7‰. Lorsque cette différenciation benthique-pélagique était non détectable, le phytoplancton (chlorophylle *a*) dominait. À ces endroits, les ratios isotopiques des brouteurs étaient similaires à ceux des filtreurs, suggérant que du phytoplancton déposé était assimilé par la communauté benthique. Nos résultats isotopiques suggèrent que les lacs fluviaux peu profonds sont constitués d'une mosaïque de secteurs allant d'une dominance complète du phytoplancton à une contribution importante du périphyton.

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## LISTE DES ABRÉVIATIONS

C	Carbone
$^{12}\text{C}$	Carbone 12
$^{13}\text{C}$	Carbone 13
chl <i>a</i>	Chlorophylle <i>a</i>
DIC	Carbone inorganique dissous
$K_d$	Coefficient d'atténuation des rayons photosynthétiques actifs
TN	Nitrates et nitrites
TP	Phosphore total
Turb	Turbidité
LSP	lac Saint-Pierre
PPR	Ratio Phytoplancton Périphyton
Phyto	Phytoplancton
Peri	Périphyton
$\delta^{13}\text{C}$	Ratio des isotopes stables du carbone
$\Delta\delta^{13}\text{C}$	$\delta^{13}\text{C}$ des brouteurs – $\delta^{13}\text{C}$ des filtreurs

## CHAPITRE I

### INTRODUCTION

À l'état naturel, les écosystèmes fluviaux sont composés d'une variété d'habitats passant de conditions lotiques, au niveau du chenal principal, à des conditions lenticques, en bordure des rives. Selon Thorp et Delong (1994), la principale source de carbone de ces écosystèmes serait de type autochtone et proviendrait des zones littorales. Plus particulièrement, les zones de faibles vitesses inondées en permanence, possèdent un potentiel de production primaire élevé et constant (Thorp et Delong, 2002). Cependant, cette production primaire, plus particulièrement la biomasse de phytoplancton (algues en suspension dans la colonne d'eau), de macrophytes (plantes aquatiques) et d'épiphytes (algues attachées à des macrophytes) peut varier de façon importante spatialement et temporellement à l'intérieur de ces zones (Vis et al., 2007). Ces variations peuvent provoquer des modifications dans la contribution relative des différents producteurs primaires au réseau alimentaire et ainsi influencer le fonctionnement de tout l'écosystème : transfert d'énergie, recyclage des nutriments, structure de la chaîne alimentaire et des habitats, etc. (Wetzel, 2001).

Les isotopes stables du carbone ( $\delta^{13}\text{C}$ ) constituent un traceur naturel de plus en plus utilisé pour évaluer l'importance relative du phytoplancton et du périphyton (algues attachées à un substrat) pour les consommateurs (invertébrés, poissons, etc.) (Forsberg et al., 1993, Lewis et al., 2001, Delong et Thorp, 2006). Les différences d'habitats entre ces deux producteurs primaires mènent à une différenciation isotopique du C d'environ 7‰ en lac profond (France, 1995b). Cependant, dans les zones de faibles vitesses des rivières, cette différenciation isotopique semble plus incertaine (Delong et Thorp, 2006). En comparaison avec les lacs profonds, ces écosystèmes possèdent généralement une colonne d'eau non stratifiée, une turbidité élevée et des patrons de mélange complexes occasionnés par l'importance des macrophytes et l'apport des nombreux tributaires. Ces

caractéristiques peuvent influencer la distribution spatiale du périphyton et du phytoplancton ainsi que leur signature isotopique.

L'objectif de cette étude est de vérifier si les isotopes stables du carbone peuvent être utilisés afin de distinguer les chaînes alimentaires phytoplanctonique et périphytique à l'intérieur des zones de faibles vitesses des rivières. Dans le cas où il y aurait chevauchement isotopique entre ces deux chaînes alimentaires, nous tenterons de déterminer quel est le maillon de la chaîne (producteur primaire ou consommateur primaire) qui est à l'origine de ce chevauchement. Chez les producteurs primaires, c'est principalement la demande en C et la limitation de ce dernier qui permettent de différencier isotopiquement le phytoplancton et le périphyton. Une variation de la demande en carbone, occasionnée par exemple par une diminution de l'intensité lumineuse chez les algues périphytiques, pourrait donc augmenter le fractionnement isotopique et ainsi mener à un chevauchement du  $\delta^{13}\text{C}$  des algues planctoniques et benthiques. Puisque le  $\delta^{13}\text{C}$  est un isotope de type conservateur, la signature isotopique des consommateurs est le reflet du  $\delta^{13}\text{C}$  de la nourriture qu'ils ont assimilée. L'ingestion simultanée d'algues planctoniques et périphytiques par les consommateurs primaires, pourrait, elle aussi, mener à un chevauchement isotopique des chaînes alimentaires phytoplanctonique et périphytique.

## RÉSULTATS

Il existe un important chevauchement entre les signatures isotopiques des consommateurs primaires filtreurs (se nourrissant de phytoplancton en suspension) et brouteurs (se nourrissant de périphyton) des zones de faibles vitesses du lac Saint-Pierre (LSP) (dernier lac fluvial du fleuve Saint-Laurent) comparativement aux résultats de France (1995b) obtenus en lacs profonds. Les résultats pairés, par station et par date d'échantillonnage des brouteurs et des filtreurs, montrent que les brouteurs sont généralement enrichis en  $^{13}\text{C}$  par rapport aux filtreurs, mais que cet enrichissement varie de 1 à 7‰.

Une expérience effectuée en milieu semi-contrôlé montre que le périphyton croissant sous de fortes intensités lumineuses (31% de la lumière incidente) possède des concentrations plus élevées en Chlorophylle *a* (Chl *a*) que celui croissant sous de plus faibles intensités lumineuses (4, 6 et 15% de la lumière incidente) (ANOVA :  $F = 34,9$  ;  $p < 0,001$ ). Ces résultats suggèrent que les taux photosynthétiques des algues benthiques diminuent lorsqu'elles sont soumises à de faibles intensités lumineuses. Les résultats du  $\delta^{13}\text{C}$  montrent que le fractionnement isotopique du C augmente significativement sous de faibles intensités lumineuses (ANOVA :  $F = 16,6$  ;  $p = 0,001$ ). La différence entre le  $\delta^{13}\text{C}$  du périphyton soumis à de faibles ou de fortes intensités lumineuses est d'environ 1‰.

La variation du  $\Delta\delta^{13}\text{C}$  ( $\delta^{13}\text{C}$  des brouteurs –  $\delta^{13}\text{C}$  des filtreurs) est expliquée par le Ratio Phytoplancton Périphyton (PPR) ( $\text{PPR} = \log ([\text{phytoplancton mg/m}^2] / [\text{périphyton mg/m}^2])$ ) ( $r^2 = 0.80$ ;  $p < 0.001$ ). La relation inverse entre le  $\Delta\delta^{13}\text{C}$  et le PPR suggère que lorsque le périphyton est abondant, les brouteurs et les filtreurs possèdent des signatures isotopiques similaires, tandis que lorsque le phytoplancton est abondant, les brouteurs et les filtreurs possèdent des  $\delta^{13}\text{C}$  similaire.

## DISCUSSION

Nos résultats isotopiques montrent un chevauchement du  $\delta^{13}\text{C}$  des filtreurs et des brouteurs et d'importantes variations du  $\Delta\delta^{13}\text{C}$ . Ce chevauchement n'est pas consistant avec la revue de littérature de France (1995b), qui montre une différenciation isotopique de 7‰ entre des organismes brouteurs et filtreurs de lacs oligotrophes profonds.

Le pourcentage de lumière incidente atteignant le fond varie entre 0,2 et 37,4% à l'intérieur de nos 12 stations au LSP. Or, les résultats de notre expérience en milieu semi-contrôlé, montrent que le périphyton croissant sous de fortes intensités lumineuses (31% de la lumière incidente) est enrichi en  $^{13}\text{C}$  comparativement au périphyton croissant sous de plus faibles intensités lumineuses (4, 6 et 15%). Ceci s'explique par

l'importance de la couche limite chez les algues benthiques qui limite la diffusion du C et, puisque le  $^{12}\text{C}$  est préférentiellement assimilé par les algues (Smith et Walker, 1980), plus la demande en C du périphyton est élevée, plus il s'enrichi en  $^{13}\text{C}$ . Cependant, une variabilité isotopique de 1‰ du périphyton, liée à l'intensité lumineuse, est insuffisante pour expliquer un chevauchement du  $\delta^{13}\text{C}$  des consommateurs primaires.

L'utilisation du  $\delta^{13}\text{C}$  des consommateurs primaires en tant qu'indicateurs des algues planctoniques et benthiques s'appuie sur le fait que ces organismes consomment uniquement des algues planctoniques dans le cas des filtreurs, et du périphyton, dans le cas des brouteurs. Or, la relation inverse observée entre le  $\Delta\delta^{13}\text{C}$  et le PPR suggère que dans un environnement dominé par le périphyton, les consommateurs primaires se nourrissent sur leur source respective de C, tandis que dans un environnement dominé par le phytoplancton, les brouteurs et les filtreurs se nourrissent de phytoplancton. Ce changement d'alimentation des brouteurs allant du périphyton au phytoplancton, selon l'abondance des sources peut être le résultat de la sédimentation du phytoplancton, lorsqu'il abonde, le rendant ainsi disponible au mode d'alimentation des brouteurs. On estime qu'au lac Saint-Pierre les algues benthiques représentent entre 96 et 27% du carbone des organismes brouteurs.

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## CHAPITRE II

### **Isotopic evaluation of the relative importance of planktonic and periphytic production for primary consumers in a large river's slackwater**

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## 15    **ABSTRACT**

16

17            Stable carbon isotopes are a natural tracer which is increasingly used to estimate  
18    the relative contribution of phytoplankton and periphyton to secondary production. This  
19    approach, developed mostly through studies of deep stratified lakes, remains to be  
20    evaluated in the context of slackwaters in large rivers. These slackwaters have a short  
21    residence time, little stratification, variable turbidity, and complex patterns of mixing  
22    due to inputs from tributaries. We used filterers and grazers as integrators of isotopic  
23    signals of phytoplankton and periphyton, respectively, at 12 stations in a fluvial lake of  
24    the Saint-Lawrence River, Canada. We found strong differences in  $\Delta\delta^{13}\text{C}$  between  
25    grazers and filterers ranging from 1 to 7‰. Areas of the lake where benthic-pelagic  
26    isotopic differentiation was not detected were dominated by phytoplankton (chl *a*). In  
27    these areas, isotopic ratios of grazers were similar to those of filterers, suggesting that  
28    deposited phytoplankton were consumed by the benthic community. Our isotopic data  
29    suggest that large river slackwaters comprise a mosaic of areas in which trophic states  
30    range from complete dominance of phytoplankton to strong reliance on periphytic  
31    production.

## 32 INTRODUCTION

33           The floodplains of large rivers comprise a variety of habitats ranging from lotic  
34 to lentic conditions. The riverine productivity model, postulates that the main source of  
35 carbon in large rivers is autochthonous primary production in the littoral zone (Thorp et  
36 Delong, 1994). Permanently inundated slackwater zones, such as shorelines,  
37 embayments and other areas outside the main channel where current velocities are  
38 substantially reduced (Thorp et Mantovani, 2005), are potentially one of the most  
39 persistent sources of autochthonous carbon for large rivers food webs (Thorp et Delong,  
40 2002). In these ecosystems, phytoplankton, periphyton and macrophytes compete for  
41 light and nutrients and their relative contribution to biomass will depend on river  
42 morphometry, depth, and nutrients (Sand-Jensen et Borum, 1991).

43           In a study on lake Saint-Pierre (LSP), a large fluvial lake with an important  
44 slackwater zone, Vis et al. (2007) used an empirical model to estimate the relative  
45 contribution of phytoplankton, epiphyton and macrophytes to total primary production.  
46 They reported important temporal and spatial variations in the relative biomasses of  
47 primary producers. These variations may lead to changes in the relative abundance and  
48 in the community structure of planktonic and benthic primary consumers. This, in turn,  
49 may influence the entire ecosystem's metabolism, the energy flow, the recycling of  
50 nutrients and, the food web and habitat structure etc. (Wetzel, 2001). For example, in a  
51 long term study of a coastal area, Josefon et al. (1993) showed that changes in the  
52 benthic community structure were influenced by increased sedimentation of  
53 phytoplankton induced by the eutrophication of the environment.

54           Stable carbon isotopes ( $\delta^{13}\text{C}$ ) are a natural tracer, increasingly used to estimate  
55 the relative contribution of periphyton (attached algae) vs phytoplankton (suspended

algae) to secondary production (invertebrates, fish) in rivers (Forsberg, et al., 1993, Lewis, et al., 2001, Delong et Thorp, 2006). In other ecosystems, such as deep lakes and marine coastal areas, the  $\delta^{13}\text{C}$  of periphyton is enriched by approximately 7‰ compared to that of phytoplankton. This isotopic enrichment is generally ascribed to differentiation in the turbulence of their respective environments (France, 1995a). Periphytic algae grow in environments that have relatively low turbulence, which favour a thicker boundary layer. This limits the diffusion of C, increases limitation by  $^{12}\text{C}$ , which is preferentially used for photosynthetic processes, and decreases isotopic discrimination (Smith et Walker, 1980).

Also, due to respiration processes,  $\delta^{13}\text{C}$  of dissolved inorganic carbon (DIC) decreases with depth in stratified lakes. This depleted carbon can be assimilated by phytoplankton and increases the isotopic differentiation between planktonic and periphytic algae (Rau, 1978). After being fixed by photosynthesis, organic carbon keeps its isotopic properties which are transferred to higher trophic levels with an enrichment lower than 1‰ per trophic transfer (Peterson et Fry, 1987).

Although this isotopic discrimination between periphytic and planktonic algae is commonly observed in stratified lakes (France, 1995b, Vander Zanden et Vadeboncoeur, 2002, Sierszen et al., 2006) it appears to be more uncertain in slackwaters of large rivers (Delong et Thorp, 2006), for several reasons: 1) compared to deep lakes, large rivers typically have a well-mixed water column, which induces vertical homogenisation of  $\delta^{13}\text{C}$  of DIC; 2) large rivers have short residence times, high turbidity, and complex patterns of mixing due to their well-developed macrophyte beds and inputs from tributaries. Those characteristics affect the spatial distribution of periphyton and phytoplankton and their carbon isotopic signature. For example, a study of the upper

80 Mississippi River, showed that transported algal matter was the major carbon source  
81 assimilated by primary consumers (DeLong et Thorp, 2006). In that study, even  
82 collector-gatherers/detritivores or scrapers had isotopic signatures corresponding to that  
83 of planktonic algae. This results, which runs counter to their expectation that isotopic  
84 ratios of scrapers should be closer to those of benthic algae, could be due to benthic-  
85 pelagic coupling induced by sedimentation (DeLong et Thorp, 2006).

86         Overlap in  $\delta^{13}\text{C}$  of food sources of filterers and grazers could also be explained  
87 by variable fractionation in periphyton as a function of light levels. High densities of  
88 phytoplankton decrease the quantity of light available for photosynthesis by periphytic  
89 algae. This causes a reduction in the C demand and, results in an increase in isotopic  
90 fractionation from DIC (Laws et al., 1995). Under light stress, periphyton would acquire  
91 an isotopic ratio similar to that of phytoplankton, and a convergence in  $\delta^{13}\text{C}$  would be  
92 observed for filterers and grazers.

93         The aim of this study is to examine the sources of carbon for primary consumers  
94 in the slackwater zone of a large fluvial lake. To achieve this goal, we used the carbon  
95 isotopic ratios of grazers and filterers as integrators of the signal for periphyton and  
96 phytoplankton. We show that the isotopic differentiation between consumers is highly  
97 variable within the lake. This result could be brought about by at least two mechanisms:  
98 1) variable fractionation by primary producers from their carbon source (s) and transfer  
99 of this signal to their consumers and 2) pelagic-benthic coupling leading to the mixing of  
100 food sources. Following an experimental approach, we first showed that the light  
101 regime, known for influencing isotopic fractionation from DIC (MacLeod et Barton,  
102 1998), accounted for very little variability in the isotopic ratio of periphyton. On the  
103 other hand, field data showed that the isotopic difference between filterers and grazers

104 was strongly related to spatial variation in the relative abundance of phytoplankton and  
105 periphyton within the lake. This study demonstrates that food web structure in the  
106 slackwater zone of a fluvial lake is spatially highly variable, ranging from reliance on  
107 distinct carbon sources to being almost solely dependent on phytoplankton.

108

## 109 **MATERIALS AND METHODS**

### 110 **Study area**

111 Lake Saint-Pierre (LSP) (mean surface area: 480 km<sup>2</sup>; mean depth: 3 m), has a  
112 large width/depth ratio, which reduces the horizontal mixing of water and, leads to  
113 persistence of three main water masses (Frenette et al., 2003). Its large littoral zone  
114 offers a variety of colonisable substrates for benthic algae and the distinctness of the  
115 water masses generates spatial heterogeneity in environmental conditions. In July and  
116 August 2006, we sampled 12 stations; 10 in slackwater area and 2 near the central water  
117 mass (Fig. 1). Each station included five sampling sites located at the four vertices and  
118 the center of a square measuring 300 m along the diagonal.

### 119 **Water characteristics**

120 At each station, we measured vertical profiles of turbidity (multiprobe: Yellow  
121 Spring Instruments, 650) and light (spectroradiometer: PUV2545, Biospherical  
122 Instruments). A sample of surface water for nutrient analyses (total nitrogen (TN) and  
123 total phosphorus (TP)) was collected (acid-washed polyethylene containers) at each  
124 station. Analyses of TN were accomplished by reduction of nitrates to nitrites by  
125 cadmium followed by spectrophotometry (APHA, 1998). Analyses of TP were  
126 accomplished by hydrolytic transformation of organic phosphorus, by persulfate and  
127 boric acid into ortho-phosphates followed by spectrophotometry (APHA, 1998).

128           We sampled each station for dissolved inorganic carbon (DIC) in 2007. Water  
129 samples collected near the surface with a syringe were immediately filtered on a 0.22µm  
130 syringe filter in an amber glass bottle to limit exchanges with the atmosphere. The bottle  
131 was completely filled and capped with a double septum, kept cold (4°C), and sent to GG  
132 Hatch Isotope Lab (University of Ottawa, Canada) the following day for C isotope  
133 analyses.

#### 134 **Primary producers**

135           Phytoplankton and periphyton biomasses at each station were determined,  
136 respectively, by the concentration of chlorophyll *a* (chl *a*) in water samples and on  
137 artificial substrates. Four litres of water from the first 1.5 m were sampled at each station  
138 for phytoplankton analyses. We introduced 1.5 m of an open PVC tube below the water  
139 and closed the top extremity in order to sample, by suction, the entire water column.  
140 Artificial substrates (10 by 10 cm porous ceramics plates) were set on June 12 and were  
141 collected between the 4 and 12 July (first sampling period), and between the 17 and 28  
142 August (second sampling period). Matter covering the artificial substrates was collected  
143 using toothbrushes.

144           In the laboratory, a homogenized fraction of each sample was collected on  
145 Millipore APFF filters (0.7µm) until the filters were visibly clogged. The filters were  
146 then frozen at -20 °C until chl *a* analyses. Filters were sonicated in cold acetone (90 %)   
147 and extraction continued in the dark for 24 hours at 4 °C. After centrifugation (5 000  
148 rpm, 5 min), we used a Turner Design fluorometer (model 10-005R) to measure chl *a*  
149 (Parsons, 1984). We averaged phytoplankton and periphyton biomasses for the 5 sites at  
150 each station/date.

151 As suggested by Vadeboncoeur et al. (2002), values of chl *a* were transformed  
 152 into  $\text{mg m}^{-2}$  by multiplying volumetric concentration ( $\text{mg chl } a \text{ m}^{-3}$ ) by depth (m). A  
 153 Phytoplankton to Periphyton Ratio (PPR) was calculated as follows:

$$154 \quad PPR = \log\left(\frac{[phyto]}{[peri]}\right)$$

155 where [phyto] and [peri] correspond to the concentration of chl *a* in  $\text{mg m}^{-2}$  for  
 156 phytoplankton and periphyton, respectively.

### 157 **Primary consumers**

158 Primary consumers are frequently used as an indicator of the  $\delta^{13}\text{C}$  of primary  
 159 producers in order to obtain values of  $\delta^{13}\text{C}$  integrated over longer time periods (Post,  
 160 2002). The difference between the  $\delta^{13}\text{C}$  of the indicator organisms of the pelagic and  
 161 benthic food web ( $\Delta = \delta^{13}\text{C}_{\text{grazers}} - \delta^{13}\text{C}_{\text{filterers}}$ ) determines if the ultimate C sources of  
 162 these two food webs are similar or not. Zoobenthos was collected at each station using a  
 163 biological dredge (mesh size of 1cm) and the contents of the dredge were kept cold in  
 164 the field until frozen in the laboratory. Grazing organisms selected as indicators of the  
 165 benthic food web were the gammarid (*Gammarus fasciatus*) and the gastropods  
 166 (*Goniobasis livescens* and *Planorbella trivolvis*). For the pelagic food web, three  
 167 bivalves (filterer organisms) were selected: *Elliptio complanata*, *Lampsilis radiata*, and  
 168 *Dressenia polymorpha*. We also selected *Bithynia tentacula* and *Viviparus georgianus*.  
 169 These gastropods are able to feed simultaneously on benthic and planktonic algae  
 170 (Brendelberger et Jurgens, 1993, Declerck, 1995).

171 For *G. Gammarus*, whole individuals were used, while only the soft body of the  
 172 gastropods and the posterior delivery muscle of the bivalves were used. Thereafter, the  
 173 samples were dried (3 days at 60 °C), crushed, then acidified drop by drop (HCl 1mol L<sup>-1</sup>



174 <sup>1</sup>) to remove carbonates (Jacob et al., 2005). The samples were dried again (three days at  
 175 60 °C) and 0.20 ±0.02mg were weighed in a tin cup for isotopes analyses. Stable  
 176 isotopes of C were analyzed with a Finnigan Delta mass spectrometer at the Stable  
 177 Isotopes in Nature Laboratory (SINLAB) (New Brunswick, Canada). Isotopic ratios are  
 178 expressed in the usual  $\delta$  notation, the deviation in ‰ being compared to a reference  
 179 standard, Pee Dee Belemnite:

$$180 \quad \delta^{13}C = \left[ \left( R_{sample} / R_{standard} \right) - 1 \right] \cdot 1000;$$

$$R = {}^{13}C / {}^{12}C$$

### 181 **Isotopic fractionation of benthic algae**

182         An experiment in a semi-controlled environment was carried out to examine the  
 183 relationship between periphyton  $\delta^{13}C$  fractionation and light intensity. During summer  
 184 2007, we installed four containers, each containing four artificial substrates (porous  
 185 ceramics plates ; 10 by 10 cm) on a floating platform on Lake Joseph (Canada), an  
 186 eutrophic lake (Simoneau et al., 2004). A pump constantly supplied lake water to the  
 187 containers, maintaining the substrates always covered by 5cm of water. Above each  
 188 substrate, the container lids were perforated and covered with screen filters letting light  
 189 pass at 4,6,15 and 31% of the incidental light. The choice of filters corresponded to  
 190 incident light levels at the bottom of the 12 stations in LSP.

191         Periphyton was collected as described above after 1 month of growth. A portion  
 192 of each sample was filtered and immediately analysed for chl *a*. The other portion was  
 193 reserved for density fractionation in order to separate the algal and detritus components  
 194 (Hamilton et al., 2005). The algal fraction was then collected on filters (Millipore APFF,  
 195 0.7µm) that had been rinsed beforehand with acid (HCl 0.1N) and pre-dried (230 °C,  
 196 during six hours). The filters were then frozen (-20 °C). Before isotopic analyses, the

197 filters were rinsed with acid ( $\text{HCl } 1\text{mol L}^{-1}$ ) in order to eliminate carbonates and then  
 198 dried following Jacob et al. (2005).

## 199 **Statistical analyses**

200  $\delta^{13}\text{C}$  of filterers and grazers species were analyzed using a paired *t*-test  
 201 procedure with a Bonferonni adjustment in order to evaluate differences between  
 202 isotopic values of species. The semi-controlled experiment results (chl *a* and  $\delta^{13}\text{C}$ ) were  
 203 analyzed using the ANOVA procedure with a Tukey post-hoc test. An ANCOVA was  
 204 used to evaluate the homogeneity of slope and y-intercept between the two regressions  
 205 linking the  $\Delta\delta^{13}\text{C}$  (difference between  $\delta^{13}\text{C}$  of filterers and grazers) and PPR. We use  $p <$   
 206 0.05 as the significance criterion for all our statistical analyses.

207

## 208 **RESULTS**

### 209 **Limnological characteristics of stations**

210 The 12 stations were highly variable with respect to physical and chemical  
 211 characteristics (Table 1). The turbidity for the two sampling periods varied between 0.3  
 212 at station 4 and 121.3 NTU at station 8 and the percentage of the incident light reaching  
 213 the bottom at each station varied accordingly between 0.2 to 37.4%. The concentration  
 214 of nutrients also varied between the 12 stations (TN ranged from 0.05 to 1.49mg/L and  
 215 TP ranged from 7.4 to 165.4  $\mu\text{gP/L}$ ). In 2007,  $\delta^{13}\text{C}$ -DIC varied between -1.6 (11) to -  
 216 12.9‰ (8) among stations (Table 1).

### 217 **Selection of isotopic integrators**

218 There were no differences between the  $\delta^{13}\text{C}$  of the filterers (mean pairwise  
 219 differences ranging from 0.16 to 0.95;  $p > 0.7$  in all cases), indicating that these three

species are consistent integrators of sestonic sources of C. Here, we report average  $\delta^{13}\text{C}$  values of filterers present at each station/date.

$\delta^{13}\text{C}$  of *B. tentacula* and *V. georgianus* were significantly lower than the other grazer species collected at the same station (mean paired differences ranging from -1.8 to -7.1;  $p = 0.016$  and  $0.014$  respectively). This isotopic shift confirms the capacity of these organisms to feed on suspended algae (Brendelberger et Jurgens, 1993, Declerck, 1995). We therefore excluded *B. tentacula* and *V. georgianus* as indicators of the periphytonic carbon source.

Mean differences in  $\delta^{13}\text{C}$  among *G. fasciatus*, *G. livescens*, and *P. trivolvus* ranged from 0.6 to 2.7 ( $p > 0.8$ ). Thus,  $\delta^{13}\text{C}$  values presented here, as indicators of the benthic food web, are the average of  $\delta^{13}\text{C}$  of *G. fasciatus*, *G. livescens*, and *P. trivolvus* found at each station/date.

### Isotopic differentiation between pelagic and benthic primary consumers

In LSP,  $\delta^{13}\text{C}$  was extremely variable for grazers and filterers, ranging respectively between -28 to -16‰ and -32 to -19‰, and therefore resulting in a substantial overlap between the  $\delta^{13}\text{C}$  of filterers and grazers (Fig. 2 A) compared to the values reported by France (1995b) (Fig. 2 B). However, averages by station/date revealed a significant correlation between  $\delta^{13}\text{C}$  of grazers and filterers ( $p < 0.001$ ). The slope of that relationship was not significantly different from one ( $t = 1.88$ ,  $df = 17$ ,  $p > 0.05$ ). Grazers were generally enriched in  $^{13}\text{C}$  compared to filterers. However, this enrichment was variable ranging from 1 to 7‰ (Fig. 3).

### Isotopic fractionation of benthic algae

Periphyton grown under high light intensity (31% of incident light) showed the highest concentration biomass ( $F = 34.94$ ;  $p < 0.001$ ) (Fig. 4 A) and isotopic

fractionation significantly increased by about one 1‰ under low light intensity (4, 6 and 15%) ( $F = 16.75$ ;  $p = 0.001$ ), resulting in more negative  $\delta^{13}\text{C}$  periphyton (Fig. 4 B).

## Contributions of phytoplankton and periphyton to primary consumers

Phytoplankton biomasses varied by approximately an order of magnitude (1.7 to 12.4  $\mu\text{g L}^{-1}$ ) while periphyton varied by more than two orders of magnitude (0.03 to 8.2  $\text{mg m}^{-2}$ ) (Table 1). The Phytoplankton to Periphyton Ratio (PPR) varied from -1.1 to 2.4 (Fig. 5). Therefore, the biomass of phytoplankton relative to that of periphyton also varied by more than two orders of magnitude among our stations. Variation in  $\Delta\delta^{13}\text{C}$  (difference between  $\delta^{13}\text{C}$  of grazers and filterers) was related to PPR ( $r^2 = 0.80$ ;  $p < 0.001$ ) (Fig. 5). Separate analyses of the two species known to be facultative grazers/filterers (*B. tentacula* and *V. georgianus*) resulted in a similar negative slopes between  $\Delta\delta^{13}\text{C}$  and PPR (ANCOVA test for homogeneity of slope;  $F = 0.655$ ;  $p > 0.05$ ), but with a significantly lower intercept (ANCOVA;  $F = 43.278$ ;  $p < 0.001$ ).

## DISCUSSION

Chlorophyll *a* results demonstrate a strong gradient in the concentration of primary producers within LSP. Phytoplankton values varied from 1.7 to 12.5  $\mu\text{g L}^{-1}$  (stations 4 and 9 respectively; Table 1) indicating that, in terms of chl *a*, some areas would be considered as oligotrophic and others as eutrophic (Wetzel, 2001). However, those variations are small relative to the large range of phytoplankton biomass reported for rivers (<1 to >400  $\mu\text{g chl a L}^{-1}$ ; Wehr et Descy, 1998). Even if periphyton concentration ranged over two orders of magnitude, our values were relatively low in comparison with stream with high nutrient loadings, maximum periphyton biomass can range from 300 to 400  $\text{mg m}^{-2}$  (Stevenson et al., 1996). The variations in density of

primary producers induced a large PPR gradient, ranging from -1.1 to 2.4, within a single ecosystem. This result indicates that in some areas, periphyton is approximately three times more abundant than phytoplankton whereas others, phytoplankton is approximately eleven times more abundant than periphyton. Even if biomass and productivity are different measure, the variations in the relative abundance of this study are comparable to those observed between phytoplankton and periphyton productivity compiled by Vadeboncoeur et al. (2002), ranging from periphyton which is about five times more productive than phytoplankton to phytoplankton which is about forty times more productive than periphyton.

A study conducted by Fry (2002), in a  $\delta^{13}\text{C}$ -DIC gradient ranging from -2 to -10‰, showed a relationship between the  $\delta^{13}\text{C}$  of the DIC and that of bivalves. Thus, the great range of primary consumers  $\delta^{13}\text{C}$  reported in LSP could be related to spatial variability in the  $\delta^{13}\text{C}$ -DIC available to primary producers. The variation in  $\delta^{13}\text{C}$ -DIC observed in LSP (-1.6 to -12.9‰; Table 1) could be explained by the inputs of low  $\delta^{13}\text{C}$ -DIC tributaries along the north and south shores of LSP, in contrast with the high carbonate, high  $\delta^{13}\text{C}$ -DIC waters arriving from the Great Lakes (Yang et al., 1996, Barth et Veizer, 1999).

Isotopic values of primary consumers showed a substantial overlap between carbon signatures of invertebrates feeding on phytoplankton and periphyton (Fig. 2) and substantial variability in  $\Delta\delta^{13}\text{C}$  (Fig. 5). This variation, in the scope of isotopic differences between grazers and filterers, is in contrast with the comparative data summarized by France (1995a) (Fig. 2 B). In that study, a very small overlap in  $\delta^{13}\text{C}$  was observed between filterers and grazers and the range of differences between these two functional groups averaged 7‰. However, these results were based on isotopic

292 ratios obtained from deep oligotrophic lakes which are subject to vertical variation in the  
 293  $\delta^{13}\text{C}$ -DIC (Rau, 1978), thus increasing the likelihood of isotopic differentiation between  
 294 phytoplankton and periphyton. In the absence of strong stratification in shallow lakes  
 295 and rivers, both phytoplankton and periphyton should be using the same pool of DIC.  
 296 Therefore, vertical variation in  $\delta^{13}\text{C}$ -DIC in slackwaters should not be a driving factor  
 297 controlling the isotope differentiation between primary producers. However, the  
 298 fractionation effects by boundary layer phenomena observed in periphyton should be  
 299 present in shallow water bodies as well as in deeper one, potentially explaining the site  
 300 specific shift between  $\delta^{13}\text{C}$  of grazers and filterers observed in the present study (Fig. 3).

301         The variation in light intensity reaching the bottom in our study (0.2 to 37.4%;  
 302 Table 1) caused a significant but modest fractionation of periphyton from DIC. The  
 303 potential fractionation linked to light levels (1‰, Fig. 4B) is insufficient to explain the  
 304 among-station variation in  $\Delta\delta^{13}\text{C}$ . In a field experiment under different light regime (100  
 305 and 10% of incident light), the  $\delta^{13}\text{C}$  of periphyton varied about 3‰ during the growth  
 306 season (MacLeod et Barton, 1998). However, that experimental light regime varied  
 307 more widely than the light conditions observed at the bottom of LSP in our experiment.  
 308 Therefore, fractionation by periphyton related to variable light regime is not likely to be  
 309 an important factor explaining variation in the isotopic ratios of grazers in LSP.

310         Another possible mechanism explaining the variability of  $\Delta\delta^{13}\text{C}$  between primary  
 311 consumers is the mixing of C sources available to filterers and grazers. The negative  
 312 relationship between  $\Delta\delta^{13}\text{C}$  and the PPR indicated that when periphyton was dominant,  
 313 grazers and filterers had distinct  $\delta^{13}\text{C}$  values and conversely, when phytoplankton was  
 314 dominant, grazers and filterers had similar isotopic signatures (Fig. 5). This suggests that  
 315 in an environment dominated by periphyton, grazers and filterers fed on their respective

316 C source. However, in an environment dominated by phytoplankton, both grazers and  
 317 filterers depended almost entirely on phytoplankton, result possibly explained by a  
 318 pelagic-benthic coupling, in which phytoplankton deposited at the bottom becomes  
 319 readily available to grazers.

320 Vadeboncoeur et al. (2003) obtained similar results in shallow productive lakes  
 321 where benthic primary consumers had  $\delta^{13}\text{C}$  close to phytoplankton primary producers.  
 322 They suggested that dominance of planktonic algae by biomass caused a diet change in  
 323 grazers, which shifted from periphyton to phytoplankton. Such changes in trophic  
 324 relationships could be related to grazing on deposited phytoplankton or to modifications  
 325 of grazer feeding systems. Our results suggest that both mechanisms could be at work in  
 326 slackwaters of fluvial lakes. Indeed, the relationship between  $\Delta\delta^{13}\text{C}$  of grazers that can  
 327 modify their feeding system to consume suspended planktonic algae (i.e., facultative  
 328 grazers such as *B. tentacula* and *V. georgianus*) and PPR is similar to that of obligate  
 329 grazers (Fig. 5). However, for a same PPR, values of  $\Delta\delta^{13}\text{C}$  of grazers/filterers are  
 330 always smaller than those of grazers. This indicates that even in an environment  
 331 dominated by periphyton, this facultative grazer uses filter feeding to collect suspended  
 332 phytoplankton. This result agrees with the study of Tashiro and Colman (1982) which  
 333 showed that *Bithynia tentacula* had greater net energy gain when they fed on suspended  
 334 phytoplankton; this adaptation may provide a competitive advantage relative to other  
 335 benthic invertebrates. Therefore, facultative grazers should not be used as indicators of  
 336 benthic food webs.

337 To calculate the relative importance of phytoplankton and periphyton to primary  
 338 consumers we assumed that: 1) filterers fed exclusively on phytoplankton, and 2)  
 339  $\delta^{13}\text{C}_{\text{periphyton}} - \delta^{13}\text{C}_{\text{phytoplankton}}$  was equal to 7‰. The first assumption is based on the

340 observation that filter feeders can only rely on planktonic algae and, therefore,  
 341 periphyton should not be an available carbon source. The second assumption seems  
 342 feasible at the light levels measured from our experimental and field results, which  
 343 suggest that there is negligible variation in the isotopic fractionation of primary  
 344 producers and that the greatest  $\Delta\delta^{13}\text{C}$  between filterers and grazers was of 7‰. Thus, the  
 345 grazers' reliance on periphytic carbon is calculated as follows:

$$346 \quad \% \text{ dependence on periphyton for grazers} = \frac{\Delta\delta^{13}\text{C}}{7\text{‰}} \cdot 100$$

347 The percent reliance of grazers on periphyton is also variable, ranging from 27% (station  
 348 9) to 96% (station 4; Table 1). The mean for the 12 station/dates shows that grazers  
 349 obtain approximately 65% of their carbon from periphyton and 35% from  
 350 phytoplankton.

351

352 Despite the marked variability of  $\Delta\delta^{13}\text{C}$  in fluvial slackwaters, our results suggest that  
 353 stable isotopic data may be used to trace the relative contribution of phytoplankton and  
 354 periphyton to consumers in these systems. Large river slackwaters comprise mosaic of  
 355 areas in which carbon flows to consumers range from dominance by phytoplanktonic  
 356 sources to strong reliance on periphytic production. Grazers can vary drastically in their  
 357 use of carbon sources, from almost sole reliance on periphyton (96%) to strong  
 358 dependence on phytoplankton (73%). Therefore, slackwater environments, because of  
 359 their shallow depth, low current velocities, and abundant macrophyte beds, are subject to  
 360 benthic-pelagic coupling, as illustrated in the present study by the deposition and  
 361 consumption of phytoplankton by benthic primary consumers.



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483  
484

485 **TABLE**

486

487 **Table 1 Averages of July and August limnological characteristics for 12 stations in LSP: depth,**  
 488 **turbidity (Turb), coefficient of light attenuation of photosynthetic available radiation ( $K_d$ ), sum of**  
 489 **nitrate and nitrite (TN), total phosphorus (TP), concentration of chl *a* in seston (Phyto),**  
 490 **concentration of chl *a* collected on artificial substrates (Peri), C isotope values of the total dissolved**  
 491 **inorganic carbon ( $\delta^{13}\text{C}$ -DIC) and the percentage of grazer carbon provided by periphytic algae**  
 492 **(graze reliance on periphyton).**

Stn	Depth m	Turb NTU	$K_d$	TN mg/L	TP $\mu\text{gP/L}$	Phyto $\mu\text{g L}^{-1}$	Peri $\text{mg m}^{-2}$	$\delta^{13}\text{C}$ -DIC (‰)*	Grazer reliance on periphyton (%)
1	1.3	37.6	4.17	1.49	18.6	4.6	2.0	-5.6	65.8
2	0.8	4.0	2.87	0.39	30.4	2.9	8.2	-8.5	94.9
3	0.9	1.1	1.78	0.21	20.7	3.6	4.9	-8.7	93.9
4	0.8	0.3	2.25	0.21	99.7	1.7	3.4	-8.6	95.9
5	0.9	1.2	2.34	0.18	11.3	1.7	6.9	.	73.5
6	0.6	30.4	1.56	0.05	26.3	3.7	1.6	-9.0	71.8
7	1.1	1.6	1.98	0.20	26.6	6.3	0.04	-7.4	38.9
8	0.6	121.3	9.76	0.88	165.4	4.2	0.5	-12.9	48.3
9	0.8	27.6	3.34	0.40	44.2	12.5	0.03	-10.4	27.0
10	0.6	61.6	2.31	0.34	42.8	3.0	1.6	-10.8	83.0
11	1.9	15.9	1.52	0.50	7.4	4.0	.	-1.6	29.7
12	1.7	5.7	2.52	0.37	34.7	6.9	4.6	-7.9	57.5

493 \* Samples for DIC were collected in August 2007.

494 **FIGURE LEGENDS**

495

496 Figure 1 Location of the 12 sampling sites in Lake Saint-Pierre.

497

498 Figure 2 Percentage frequency distribution of  $\delta^{13}\text{C}$  (‰) for filterers and grazers in (A)  
499 the present study and (B) the study of France (1995b).

500

501 Figure 3  $\delta^{13}\text{C}$  (‰; mean and standard deviation) of grazers and filterers for each station  
502 in July and August 2006.  $\delta^{13}\text{C}_{\text{grazers}} = -2.7 + 0.7 \cdot \delta^{13}\text{C}_{\text{filterers}}$  ( $p < 0.001$ ).

503 The 1:1 line is shown.

504

505 Figure 4 Box plots of Chl *a* concentration ( $\text{mg m}^{-2}$ ) (A) and  $\delta^{13}\text{C}$  (‰) of benthic algae  
506 (B) in relation to the percentage of incident light reaching the artificial substrate during  
507 the experiment in a semi-controlled environment, letters represent the results of Tukey  
508 post-hoc test of an ANOVA analyse.

509

510 Figure 5:  $\Delta\delta^{13}\text{C}$  (‰); the difference between the  $\delta^{13}\text{C}$  of grazers (dark circle) and of  
511 grazers known to be able to filter (*B. tentacula* and *V. georgianus*) (open circle) and the  
512  $\delta^{13}\text{C}$  of filterers, in relation with the index of abundance of phytoplankton (PPR) for the  
513 12 stations at the two sampling dates.

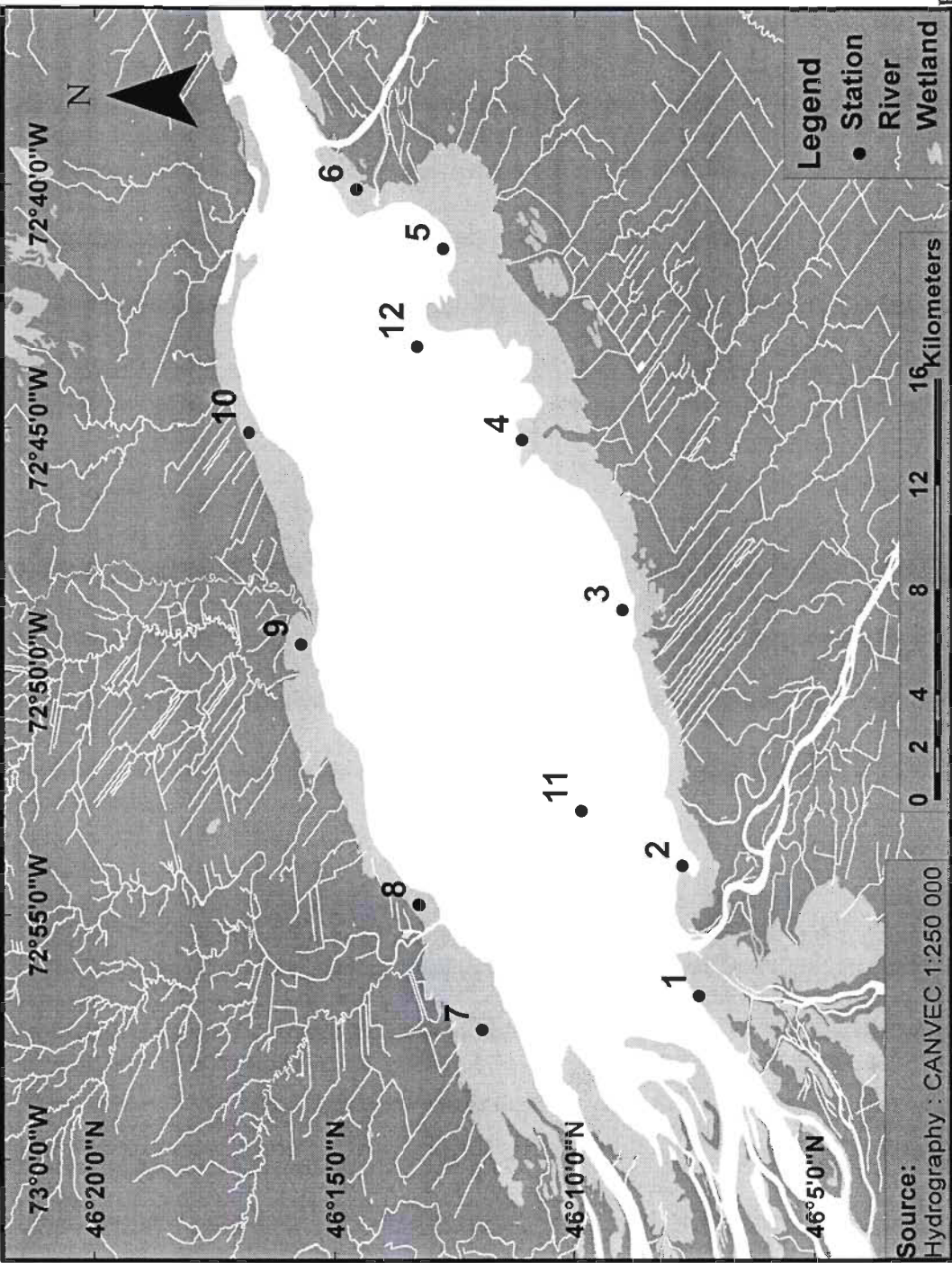
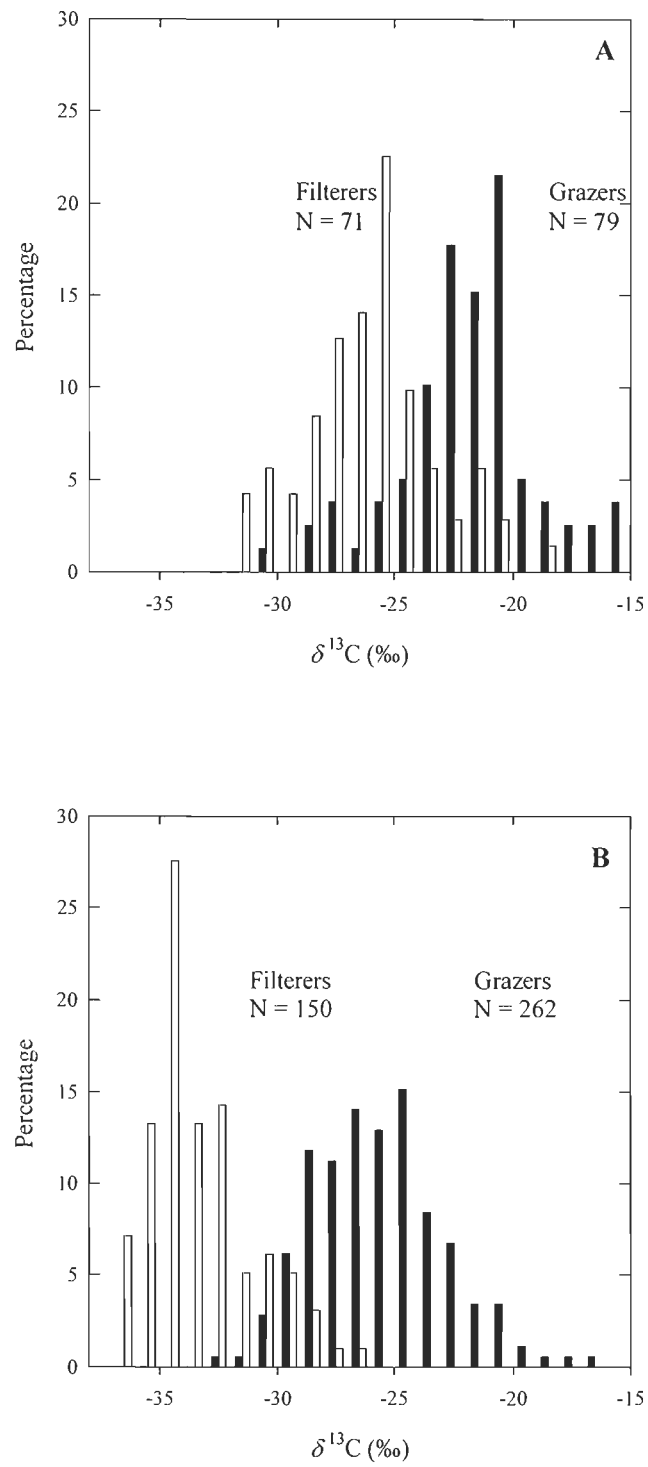


Fig. 1

**Fig. 2**

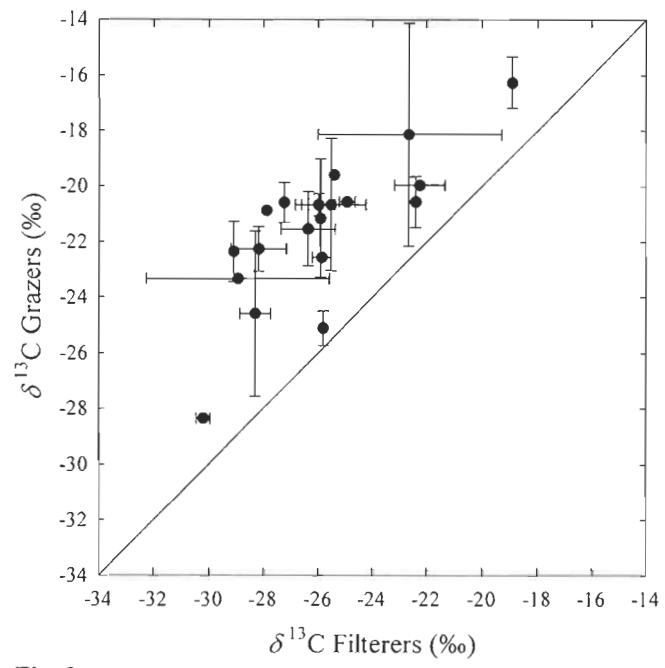


Fig. 3



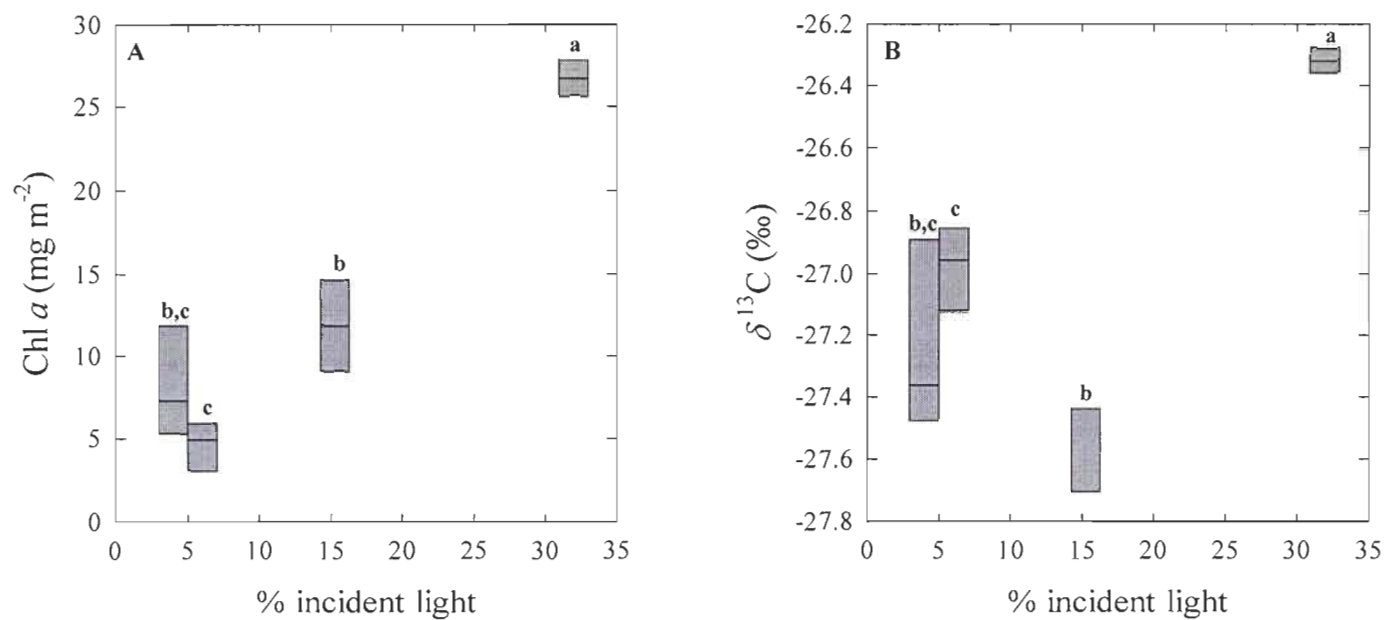


Fig. 4

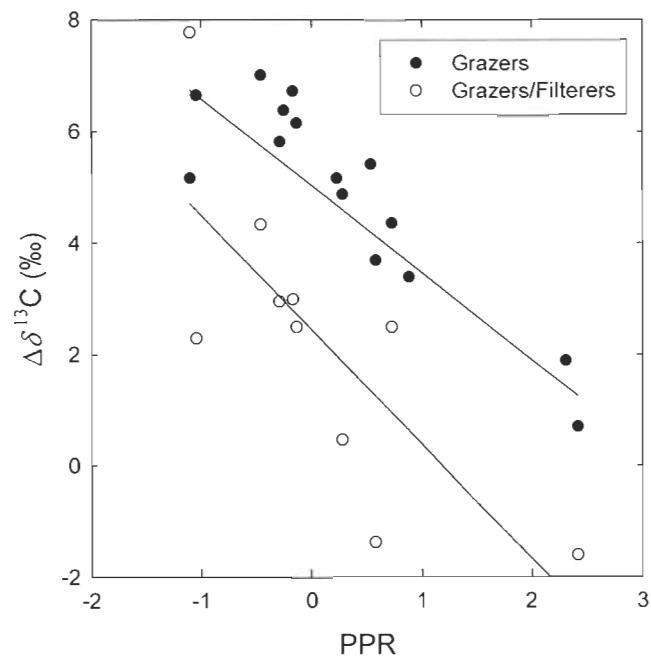


Fig. 5

## ANNEXE

### Author Instructions

#### General points

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Capone [eds.], *Nitrogen in the marine environment*. Academic.

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Use the Times New Roman font for all text and numerals on figures. Font sizes should be from 9 to 11 points. If mathematical or Greek symbols are not available in Times New Roman, use the Symbol font.

Page layout: See page layout diagram.

Submit figures at the intended print size. The L&O column width is 8.9 cm (3.5 in) and full page width is 18.4 cm (7.25 in). The maximum size for a figure is 18.4 x 23.2 cm (7.25 x 9.125 in).

Make figures as simple as possible. For example, avoid grid lines and boxes around symbol definitions.

Maps must include latitude and longitude, an indication of compass direction, and a thin line as a border. All markings must be legible.

If a figure is submitted as mounted artwork, mount it on flexible paper because it will be scanned on a drum scanner; use glue stick to attach just the top edge of each panel to the paper, making sure that the plate is flat (i.e., there are no bumps or bubbles); any unevenness will cause distortion of the final image.

Color figures:

See detailed instructions.